

Contents lists available at ScienceDirect

# Forest Ecology and Management



journal homepage: www.elsevier.com/locate/foreco

# The dynamics of conspecific tree and seedling neighbors on seedling survival in a subtropical forest



Heming Liu<sup>a</sup>, Daniel J. Johnson<sup>b</sup>, Qingsong Yang<sup>a</sup>, Mingjie Xu<sup>a</sup>, Zunping Ma<sup>c</sup>, Xiaofeng Fang<sup>d</sup>, Yijing Shang<sup>a</sup>, Xihua Wang<sup>a,\*</sup>

<sup>a</sup> Zhejiang Tiantong Forest Ecosystem National Observation and Research Station, School of Ecological and Environmental Sciences, East China Normal University, 200241, China

<sup>b</sup> School of Forest Resources and Conservation, University of Florida, Gainesville, FL 32611, USA

<sup>c</sup> Mianyang Teachers' College, Mianyang, Sichuan 621000, China

<sup>d</sup> School of Water Resources and Environment, Hebei GEO University, Shijiazhuang, Hebei 050031, China

### ARTICLE INFO

Keywords: Intraspecific negative interactions Conspecific Heterospecific Species coexistence Seedling age

### ABSTRACT

Negative intraspecific interactions could provide opportunities for heterospecific regeneration, thereby maintaining species coexistence in forest communities. Increasing conspecific tree and seedling neighbor densities often have negative correlations with seedling survival. If and how the strength of interactions change with seedling age remains a knowledge gap. In this study, we followed 2184 newly germinated seedlings belonging to 26 species for five years in a subtropical evergreen broad-leaved forest in eastern China. We tested the density distribution pattern of different aged live seedlings and used generalized linear mixed models to assess the dynamic influence of conspecific tree and seedling neighbors on seedlings survival across years. Our results showed that the density distribution of newly germinated seedlings was more clustered than the other aged live seedlings. Over five years, the clustered density distribution of live seedlings gradually weakened. Particularly, the density distribution of five years old seedlings was statistically random. Meanwhile, the degree of positive correlation between the density of live seedlings and the density of conspecific tree neighbors decreased with seedling age. Correspondingly, the conspecific seedling neighbors had a significant negative correlation with newly germinated seedlings survival in the first year and had a greater effect than conspecific tree neighbor density. As seedlings age, the conspecific tree neighbors had a greater negative relative effect on seedlings survival than the conspecific seedling neighbors. Our findings demonstrate that intraspecific negative interactions are an important mechanism on seedling survival and the magnitude of those interactions change with seedling age. In younger seedlings, intraspecific negative interactions were mainly from conspecific seedling neighbors and weakened as the seedlings aged. In later seedling ages, the interactions from conspecific tree neighbors dominated. Ultimately, these effects reduce species clumping dynamically and provide space for establishment of heterospecifics which would maintain diversity in the forest.

### 1. Introduction

Understanding the mechanisms of species coexistence in communities is an enduring ecological challenge. In recent decades, ecologists have increasingly found density-dependent predation and intraspecific resource competition would result in negative intraspecific interactions (Chesson, 2000). This process increased relative to interspecific interactions causing heterospecific species stabilizing regeneration (Tilman, 1982, Chesson, 2000, Adler et al., 2018). In addition, Janzen (1970) and Connell (1971) have found conspecific trees, as the host of specific pathogens and herbivores, could attract or transmit these natural enemies on their offspring. Janzen-Connell effects (hereafter called J-C effect) also result in negative intraspecific interaction, thus maintaining species diversity in the community.

Conspecific tree and seedling neighbors were regarded as the vital factors of intraspecific negative interactions on seedling survival (Webb and Peart, 1999, Comita et al., 2010, Chen et al., 2010, Johnson et al., 2012). Many field and laboratory studies of the J-C effect have proved

https://doi.org/10.1016/j.foreco.2021.118924

Received 5 November 2020; Received in revised form 30 December 2020; Accepted 31 December 2020 Available online 15 January 2021 0378-1127/© 2021 Elsevier B.V. All rights reserved.

<sup>\*</sup> Corresponding author at: School of Ecological and Environmental Sciences, East China Normal University, 500 Dongchuan Road, Shanghai 200241, China. *E-mail address:* xhwang@des.ecnu.edu.cn (X. Wang).

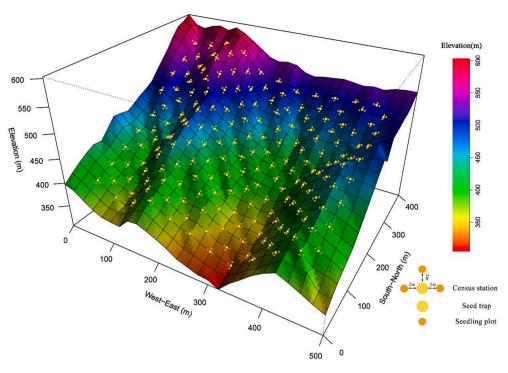


Fig. 1. Distribution of seedling plots in the Tiantong plot.

specific pathogens and herbivores is clumped in large trees, thus inhibiting conspecific seedlings survival around them (Norghauer et al., 2006, Liu et al., 2012, Bagchi et al., 2014, Liu et al., 2015, Chen et al., 2019). This seedling mortality would decrease with the distance to the conspecific trees (Terborgh, 2012, Comita et al., 2014, Murphy et al., 2017). While, conspecific seedling neighbors clumping could also increase seedling mortality due to rapid spread of natural enemies (Carson et al., 2008, Bagchi et al., 2014). In addition, seedlings aggregating to the conspecific trees (Janzen, 1970) would also cause intraspecific asymmetric (tree-seedling) (Tomé et al., 1994) and symmetric (seedlingseedling) competition (Grubb, 1977) for similar resource requirements (Hogan, 1986, Weiner and Thomas, 1986, Tomé et al., 1994, Lentz, 1999, Lewis and Tanner, 2000).

Density of conspecific seedlings determine the intensity of intraspecific negative interactions on seedling survival (Chesson, 2000). The conspecific high-density seedlings would be detected by densityresponsive predator easily (Connell, 1971) or suffer the symmetric competition (Weiner and Thomas, 1986, Ammer et al., 2010). Therefore, the intensity of intraspecific negative interactions would be increased in high density area. Correspondingly, low density of conspecific seedlings would avoid density-responsive predator and symmetric competition (Connell, 1971, Harms et al., 2004), thus decreasing intensity of intraspecific negative interactions. Whereas the effect of conspecific trees would come out through distance-responsive predation (Janzen, 1970).

As seedlings die, the density distribution of live seedling would be changed. Typical seed dispersal produce a mass of offspring around the parent tree (Forcier, 1975). The density distribution of newly germinated seedlings would aggregate to the conspecific trees (Hubbell, 2001). However, due to negative density dependent mortality, the high density of seedlings would suffer with large reductions (Janzen, 1970). As time goes on, the aggregate distribution would be weakened. In the later stage of seedlings, the effects of physical environment will determine the density distribution of seedlings (Grubb, 1977).

The density distribution of seedlings can vary with age, however, few studies focus on the temporal distribution characteristic of seedling densities. Meanwhile, this characteristic will cause changes of intraspecific interactions. Ignoring it would increase bias in detection of intraspecific interactions. Here, we examine the dynamics of seedling survival in a subtropical forest over the course of five years in relation to conspecific tree and seedling neighbors. Specifically, we calculate the coefficient of variation of seedling densities and built generalized linear mixed models of seedling survival to address the following questions: (1) How did the density distribution of seedlings change with age? (2) Does the conspecific tree and seedling neighbors have correlations with seedling survival varying with time? (3) If so, do these dynamics respond to temporal variation of seedling density distribution pattern?

# 2. Materials and methods

## 2.1. Study site and data collection

The study was conducted in a 20 ha (500 m  $\times$  400 m) dynamics plot in a subtropical broad-leaved evergreen forest in Tiantong Forest Park in eastern China (hereafter called Tiantong plot) (Yang et al., 2011). This region has a subtropical monsoon climate and receives 1374 mm mean annual rainfall. The hot-humid-summer is from June to August; colddry-winter is from December to February. Mean annual temperature is 16.2 °C, the maximum mean monthly temperature is 28.1°C and minimum is 4.2°C (Song and Wang, 1995). All trees (DBH  $\geq$  1 cm) in Tiantong plot were tagged, mapped, measured and identified to species in 2010. The dominant families are Theaceae, Lauraceae and Fagaceae (Yang et al., 2011).

To monitor seedling dynamics in Tiantong plot, we established 187 seedling census stations in two diagonal corners of each 20 m  $\times$  20 m area within the Tiantong plot. To avoid conspectic tree neighbors out of plot, we selected the stations at least 40 m far away from the border (Fig. 1). Each census station includes one seed trap in the center and three 1 m  $\times$  1 m seedling plots. These seedling plots were located at 2 m from the center of seed trap in three random directions (Fig. 1). From October 2011 to April 2012, we tagged all the free-standing woody seedlings (DBH  $\leq$  1 cm) in 561 seedling plots. In October 2012, the newly germinated standing woody seedlings in seedling plots were tagged, mapped, measured to initial height and identified to species. We investigated 2,184 newly germinated seedlings belonging to 26 species, 19 genera, 11 families. Then, we monitored these seedlings from

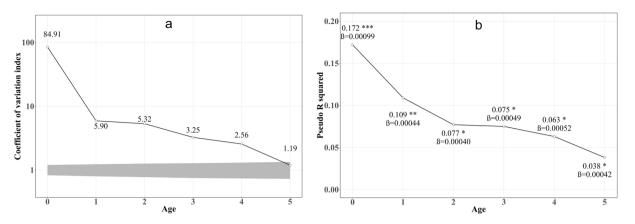


Fig. 2. Coefficient of variation index of conspecific live seedling density (a) and marginal pseudo R squared between the conspecific tree neighbors index and live seedlings density (b) in different ages. Age of "0" represent the newly germinated seedlings. Gray area represents the 95% confidence interval. Above of confidence interval represent the clumped density distribution; Below of confidence interval represent the disperse density distribution; Between of confidence interval represent the random density distribution. "\*\*\*" represent *P* value < 0.001, "\*\*" represent *P* value < 0.01, "\*" represent *P* value < 0.05. $\beta$  represent coefficient of conspecific tree neighbors index.

October 2012 to October 2017. In October of each year during the study period, we surveyed the height and survival condition of these seedlings.

## 2.2. Data analyses

To test how density patterns of live seedlings differ with age, we calculate the coefficient of variation of conspecific seedling densities. The value of the coefficient of variation will increase with cluster densities (Blackman, 1942). To combine coefficients of variation for all species and consider their importance, we multiply by the station occurrences of each species to create a weight. This index can be specified as:

$$C_l = \operatorname{Var} (x_l) / \operatorname{E} (x_l) \tag{1}$$

$$w_l = n_l \bigg/ \sum_{l=1}^m n_l \tag{2}$$

$$C = \sum_{l=1}^{m} c_l w_l \tag{3}$$

where  $C_l$  is the coefficient of variation of species *l*(except the species whose live seedlings are in only one station).  $x_l$  is the live seedling numbers of species *l*in each station (equation (1)). $w_l$  is the weight of each species *l*.  $n_l$ is the number of stations containing live seedlings of species *l*. *m* is the number of species (equation (2)). *C* is the coefficient of variation index (equation (3)).

In order to estimate density distribution pattern of live seedlings (clumped, disperse or random), we simulate the live seedling densities in Poisson distribution for 1999 times and calculate their coefficient of variation index (Crandom). Then, through Cactual comparing with Crandom, we could judge the actual density pattern of live seedlings. The detailed calculation process are as follows. Firstly, we calculate the seedling numbers  $x_l$  of each species l in different station (set of seedling densities). Then, we use mean of  $x_l$  ( $\lambda_l = \sum x_l/n_l$ ) and station numbers  $n_l$ build set of seedling densities in Poisson distribution for species l. After simulating of the whole species, we could calculate the C for this simulated data. We repeated the aboved steps for 1999 times. We could have Crandom. Finally, we calculated the 95% confidence interval of Crandom (lower limit is 50th minimum value; upper limit is 50th maximum value) to test the density distribution pattern of live seedlings. Cactual above of confidence interval of Crandom indicates the density distribution of live seedlings is clumped; Cactual below indicates dispersion; Cactual between indicates random.

To examine the correlation between the density of live seedlings and conspecific tree neighbors along with the survival time, we built generalized liner mixed-effects models (GLMMs) with Poisson errors for different ages of live seedlings. The model with the random effects can be specified as:

$$Y_{ikl} \sim \text{Poisson} \ (\lambda_{kl}) \tag{4}$$

$$Log(\lambda_{kl}) = [\alpha + \beta \times x]_{\text{fixed.part}} + [\mu_l + \mu_k]_{\text{random.part}}$$
(5)

where  $Y_{ikl}$  is number of live seedlings of species l in seedling plot i belonging to seedling station k,  $\lambda_{kl}$  is the parameter of Poisson distribution.  $\lambda_{kl}$  is equal to expected value and variance of  $Y_{ikl}$  (equation (4)). Position ( $\mu_k$ ) and species ( $\mu_l$ ) are the two subparts of the random effects. In the fixed part of the equation (5),  $\alpha$  and  $\beta$  represent an intercept and a vector of coefficients of explanatory variables x respectively. x is the index of conspecific tree neighbors which were defined as the total of the basal area of conspecific trees (DBH  $\geq 1$  cm) within best fit range (based on comparison of models using different neighborhood radii; see the details in Table S1) and divided by the distance from the focal tree to the center of focal seedling plot (Comita and Hubbell, 2009, Lin et al., 2012, Lin et al., 2014, Liu et al., 2016). We calculate the marginal pseudo-R-squared of fixed effect (Johnson, 2014) to estimate how much the distribution of conspecific tree neighbors can explain the density of live seedlings.

To examine the dynamic effect of conspecific tree and seedling neighbors on seedling survival in Tiantong plot and consider other potential biotic and abiotic factors, we built GLMMs with binomial errors in each year of study period. The model with the random effects can be specified as:

$$Y_{ijkl} \sim \text{binomial} (1, \pi_{ijkl})$$
 (6)

$$logit(\pi_{ijkl}) = [\alpha + \beta \times x]_{fixed,part} + [\mu_l + \mu_{j|k} + \mu_k]_{random,part}$$
(7)

where  $Y_{ijkl}$  is 1 if focal seedling *i* of species *l* in the seedling plotjof station *k*alive in one year and 0 otherwise,  $\pi_{ijkl}$  is the survival probability of focal seedling (equation (6)). Position  $(\mu_{j|k} + \mu_k)$  and species  $(\mu_l)$  are the two subparts of random part. Especially, the position has two levels, first level was seedling plot *j* belonging the station  $k(\mu_{j|k})$  and seconds is station k ( $\mu_k$ ). $\alpha$  and  $\beta$  is same as equation (5).

In this study, in addition to the intraspecific factors of conspecific tree and seedling neighbors, we also used heterospecific tree and seedling neighbors, and seedling height. The conspecific and heterospecific tree neighborhood index were similar with above mentioned (index of

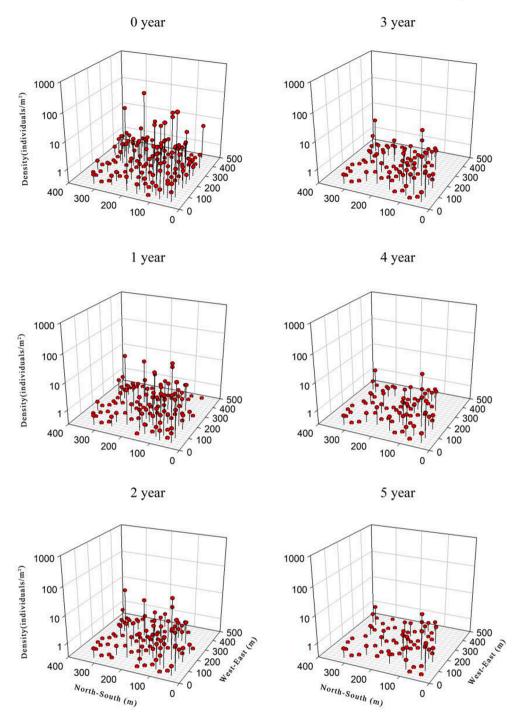


Fig. 3. Density distribution of live seedlings in different ages. Age of "0" represent the newly germinated seedlings. Z aixs represent the density of total live seedlings in a station.

conspecific tree neighbors). The best fit neighborhood is based on comparison of models in Table S2 and the distance is from the focal tree to the center of seedling plot including the focal seedling. The conspecific and heterospecific seedling neighborhood index were density of conspecific or heterospecific seedlings which were located in the same seedling plot and germinate in the same time. Seedlings germinating in the same time are likely to be influenced by symmetric seedling competition (Berkowitz et al., 1995, Davis et al., 1998, Paine et al., 2008, Peay, 2018, Wang et al., 2018, Wu et al., 2018). Meanwhile, Taylor and Aarssen (1989) found the seedling competition was from the mast seedlings in the same year. Therefore, seedling height is defined as the previous year's height. The abiotic factors include topography (elevation, aspect and slope), light (canopy openness) and soil condition (pH value, total nitrogen concentration and total phosphorus concentration). The elevation was measured by total station on center of each seedling station. The aspect and slope of each seedling plot were measured by compass and tripod with three axes of rotation (Benro A650FHD3, China, Guangdong) (details of measuring and calculating methods in Xing, 2015). In addition, aspect values were linearized into a north–south range with the formula cosine (x) + 1.1, where x is the aspect in degrees from north (Wang et al., 2007). The canopy openness of each seedling plot was estimated by the photograph 10 cm above ground level and calculated by a canopy image analysis software (WinSCANOPY; Regent

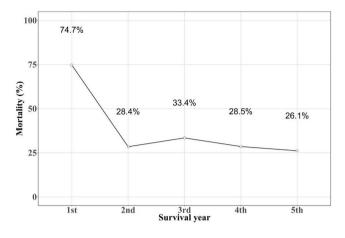


Fig. 4. The mortality of seedlings across years.

Instruments INC, Canada) in RGB mode. Soil samples were collected in the center of seedling station (details of collecting methods in Fang et al., 2017). Soil pH value (pH) were analyzed with a Beckman glass electrode. Soil total nitrogen concentration (TN) was measured by an elemental analyzer (Vario Micro Cube; Elementar, Langenselbold, DE). Soil total phosphorus concentration was measured by a continuous flow analyzer (SAN<sup>++</sup>; Skalar, Breda, NL).

We use *AICc* weight and model-average estimator of potential explanatory variable to estimate the relative importance of potential factors on seedling survival respectively (Burnham and Anderson, 2002). Firstly, we select the *m* explanatory variables to combine the fixed variables from the *n* potential explanatory variables without repeat, resulting in  $2^n (\sum_{i=0}^n C_n^i = 2^n)$  GLMMs. Secondly, we calculated the *AICc* weight of potential explanatory variable (equations S1 and S2) and estimated the *AICc* weight of potential explanatory variable (equations S3 and S4). Finally, we selected the optimal models ( $\Delta AICc \leqslant 2$  Table S3) to calculate the model-average estimator ( $\hat{\beta}$ ) and standard error ( $\hat{s.e.}$ ) (equations S5, S6, S7 and S8). The methods for calculating *AICc* weight showed the relative importance of the explanatory variables. In addition, the model-averaged variables estimate the correlations of key biotic and abiotic factors on seedling survival by synthesizing the optimal models and reducing the multi-collinearity (Liu et al., 2016).

All explanatory variables were normalized by the mean of the variable and dividing by the standard deviation. All analyses were conducted in R 3.3.2 (R Development Core Team 2018). The GLMMs were fit by the "glmer" function of "Ime4 1.1–13" package (Bates et al., 2013) and marginal pseudo-R-squared were fit by the "r.squaredGLMM" function of "MuMIn 1.40.0" packages (Nakagawa and Schielzeth, 2013, Johnson, 2014).

# 3. Results

# 3.1. The density distribution of live seedlings in different ages

The density distribution of newly germinated seedlings was most clustered of any age class (Fig. 2a). The strength of this clustered density distribution weakened with time (Fig. 2a). Furthermore, the five-year-old seedlings' density distribution did not depart from null expectations of randomness (Fig. 2a). In addition, newly germinated seedlings was most correlated with conspecific tree neighbors' aggregation (Fig. 2b). Then, this correlation also weakened with time (Fig. 2b). The maximum density of newly germinated seedlings in one station was 241 individuals/m2 (Fig. 3). Over five years, 82.3% of seedling stations include 1 or 2 five-year-old seedlings (Fig. 3).

# 3.2. Temporal dynamics of conspecific tree and seedling neighbors on seedling survival

The newly germinated seedlings suffered the highest mortality (74.7%) in the first year. In the next year, mortality decrease to 28.4% (Fig. 4). Conspecific seedling neighbor density had a significant negative correlation on newly germinated seedlings survival (Fig. 5). During the next four years, the conspecific seedling neighbors did not have significant correlation with seedling survival (Fig. 5). Conspecific tree neighbors did not have significant correlations with seedling survival during the first three years (Fig. 5). However, in the fourth and fifth survival year, significant negative correlations were observed for seedling survival with a larger relative effect size than conspecific seedling neighbors (Fig. 5). Seedling height had consistent significant positive correlations with seedling survival and had the largest relative effect size among the biotic factors in the first four years. The heterospecific tree neighbors had significant positive correlations with seedling survival in second and third survival year. (Fig. 5). Canopy openness had significant positive correlations with seedling survival in the first two years.

# 4. Discussion

The negative relations of conspecific densities on seedling survival varied with time. In our study, we found clumped density distribution of newly germinated seedlings tended to random most intensively in the first survival year and only because of conspecific seedling neighbors. As the seedlings age increasing and density distribution tending toward random, the negative relations with conspecific tree neighbors emerge as significant factors for seedling survival suggesting a switch from symmetric to asymmetric intraspecific negative interactions during early ontogeny.

Due to dispersal limitation, newly germinated seedlings were clustered nearby the conspecific trees (Willson and Traveset, 2000). Previous studies point towards natural enemies existing in high concentrations around conspecific densities. Both the conspecific seedling and tree neighbors should have significant negative effects on newly germinated seedling survival for establishment (Augspurger, 1984, Augspurger and Kelly, 1984, Packer and Clay, 2000, Comita and Hubbell, 2009, Liu et al., 2012, Jurinitz et al., 2013, Johnson et al., 2014, Johnson et al., 2017). However, in our study, we only found the conspecific seedling neighbors have significant relation to newly germinated seedlings establishment and conspecific tree neighbors' relation was the least influential biotic factor in this stage. These results indicate intraspecific negative interactions in seedling establishment was mainly related with conspecific seedlings neighbors in this study region. The conspecific tree neighbors determined seedlings distribution in early stage, but not necessarily the dynamics. The reason for this phenomenon could be newly germinated seedlings is not established yet, thus easily suffering from density-responsive predation in the area of conspecific seedling neighbors clumping (Connell, 1971, Hyatt et al., 2003). In addition, high density of conspecific seedling neighbors would cause the overlap of above- and underground resource demands (Hogan, 1986, Lentz, 1999). Thus, newly germinated seedlings would be difficult to established in the area with high density of conspecific seedling neighbors.

The distance-dependent mortality from conspecific tree neighbors became more influential on seedling dynamics in later ages (Murphy et al., 2017). After seedling established, their densities declined rapidly. Lower density would cause the symmetric intraspecific negative interactions to weaken (Paine et al., 2008). Whereas, distance-responsive predation from conspecific tree neighbors has little impact on low seedling densities (Janzen, 1970). Thus, the soil-borne species-specific pathogens would infect on established seedlings around conspecific trees (Carson et al., 2008, Liu et al., 2015). Furthermore, the older individuals would be easier to suffer from soil-borne species-specific pathogens on their root systems (Packer and Clay, 2000). Therefore, we

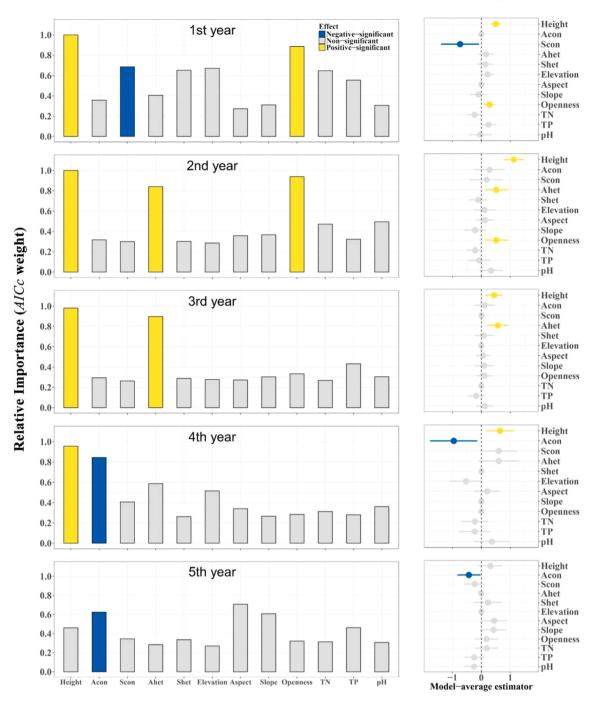


Fig. 5. The relative influence of potential factors on seedling survival across years. The variables examined are the log-transformed seedling height (HT), conspecific tree neighborhood index (Acon) and seedling density (Scon), heterospecific tree neighborhood index (Ahet) and seedling density (Shet), elevation, aspect, slope and canopy openness (Openness) in seedling plot, total nitrogen (TN), total phosphorus (TP) and pH value of soil. Error bars represent 1.96  $\widehat{s.e.j}$  around the model-average estimator ( $\widehat{\beta}_j$ ). Blue and yellow point indicate parameter estimates significantly different from zero at the alpha = 0.05 level. Some gray point with no error bars indicates this parameter estimate does not include in the optimal models. (For interpretation of the references to colour in this figure legend, the reader is referred to the web version of this article.)

could found the effect of conspecific tree neighbors on seedling survival in later ages.

Seedling ages play a crucial role on seedlings density distribution and dynamics. In general, newly germinated seedlings suffered the highest mortality in the first survival year for establishment (Gilbert et al., 2001, Murphy et al., 2017, Kuang et al., 2017). Only a fraction of seedlings could be established through this demographic bottleneck for population (Gurevitch et al., 2006). Due to symmetric intraspecific negative interactions, the mortality risk would increase with the density of

seedlings (Chesson, 2000). Therefore, the aggregated density distribution tends to random rapidly and conspecific seedling neighbors turn a dominant effect. As seedling established, seedlings mortality is stable. Seedlings turns to develop by absorbing above- and below ground resource (Grubb, 1977). In this stage, negative plant-soil feedback around conspecific tree would inhibit established seedling growth. Over longer periods of time, seedlings with slow growth rate would be likely to die (Mangan et al., 2010). Thus, in the later stage of seedlings, the distance-dependent seedling mortality from conspecific tree neighbors

## could be detected.

Although the negative intraspecific interactions are potential mechanism affecting seedling density distribution and dynamics, the environmental factors influencing survival should not be neglected, especially light condition (Augspurger, 1984, Augspurger and Kelly, 1984, Dai et al., 2002, Lin et al., 2014, Inman-Narahari et al., 2016). We found light availability had a significant positive effect on seedling survival in the early stages, indicating newly germinated seedlings benefit from well-lit conditions. Because, the high light conditions could reduce pathogen activity (Augspurger and Kelly, 1984). However, light effects weakened as seedling ages. Because, light promote established seedlings growth more efficiently than survival (Liu et al., 2017). This would be the reason that Johnson et al. (2017) could not found significant effect of light on old seedlings survival. In addition, due to the steep topography of Tiantong plot (Yang et al., 2011), the density distribution of seedlings are influenced by topography factors, especially slopes. The slope have a significant negative effect on density distribution of seedlings in all ages (Fig. S1). The reason is that most seedlings are hardly to regenerate in steep slope with bare soil in evergreen forest (Kitajima and Fenner, 2000). In addition, the heterospecific tree neighbors crowding could result in fewer encounters between a host and its species-specific pests and pathogens (Wills, 1996, Peters, 2003). Thus, the focal seedlings survival would increase with the heterospecific tree neighbors crowding. In species-level, the clumped density distribution of newly germinated seedlings was mainly belonged to Machilus leptophylla (Fig. S2). Furthermore, its conspecific seedling neighbors also have significant negative correlation with seedling establishment (Fig. S3). Therefore, we estimate the negative correlation of conspecific seedling neighbors on establishment would contribute to this prolifically germinating species. If there is no dominant species with mast newly germinated seedlings, we would hardly detect this negative correlations.

Conspecific seedling and tree neighbors were major intraspecific negative interaction factors influencing seedling survival and the relative effect varied with seedling age. Most previous studies for conspecific negative density dependence focus on the seedling stage as a whole. Our findings suggest that there are reversals in the influence of each factor as seedlings age and not accounting for this variation could lead to incomplete understand of the effects across the seedling stage and how different factors wax and wane in importance to tree community dynamics and diversity.

## 5. Authors' contributions

HL, DJJ and XW conceived and designed the ideas. HL, QY, MX, ZM, XF and YS collected the data. HL and QY analyzed the data and contributed analysis tools. HL, DJJ and XW drafted and revised the article. All authors agree to be accountable for the aspects of the work.

# **Declaration of Competing Interest**

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

# Acknowledgment

We thank Siyuan Ren, Jiuzhou Xing, Qingkai Lin, Yi Zong, Xiyang, Fei, Tongtong Zhu for help with the field work. Meanwhile, we are also grateful to Fangliang He, Guochun Shen and Yu Liu for their constructive comments on this study. This study was financially supported by the National Natural Science Foundation of China (Grant No. 31800351), and "Fundamental Research Funds for the Central Universities".

### Data Accessibility

Our data will be archived at https://github.com/kkkliuheming/seed ling-data.

# Appendix A. Supplementary data

Supplementary data to this article can be found online at https://doi.org/10.1016/j.foreco.2021.118924.

### References

- Adler, P.B., Smull, D., Beard, K.H., Choi, R.T., Furniss, T., Kulmatiski, A., Meiners, J.M., Tredennick, A.T., Veblen, K.E., 2018. Competition and coexistence in plant communities: intraspecific competition is stronger than interspecific competition. Ecol. Lett. 21, 1319–1329.
- Ammer, C., Förster, B., Höllerl, S., 2010. The impact of intraspecific competition in early stages of European beech (Fagus sylvatica L.) regeneration. Forstarchiv 81, 175–180.
- Augspurger, C.K., 1984. Seedling survival of tropical tree species: Interactions of dispersal distance, light-gaps, and pathogens. Ecology 65, 1705–1712.
- Ausspurger, C.K., Kelly, C.K., 1984. Pathogen mortality of tropical tree seedlings: experimental studies of the effects of dispersal distance, seedling density, and light conditions. Oecologia 61, 211–217.
- Bagchi, R., Gallery, R.E., Gripenberg, S., Gurr, S.J., Narayan, L., Addis, C.E., Freckleton, R.P., Lewis, O.T., 2014. Pathogens and insect herbivores drive rainforest plant diversity and composition. Nature 506, 85–88.
- Bates, D., Maechler, M., Bolker, B., Walker, S., 2013. lme4: Linear mixed-effects models using Eigen and S4. R package version 3 (2), 1. http://lme4.r-forge.r-project.org/.
- Berkowitz, A.R., Canham, C.D., Kelly, V.R., 1995. Competition vs. facilitation of tree seedling growth and survival in early successional communities. Ecology 76, 1156–1168.
- Blackman, G.E., 1942. Statistical and ecological studies in the distribution of species in plant communities: I. Dispersion as a factor in the study of changes in plant populations. Ann. Bot. 6, 351–370.
- Burnham, K.P., Anderson, D.R., 2002. Model Selection and Multimodel Inference: A Practical Information-Theoretical Approach, 2nd ed. Springer, New York.
- Carson, W.P., Anderson, J.T., Leigh, E.G., Schnitzer, S.A., 2008. Challenges associated with testing and falsifying the Janzen-Connell hypothesis: a review and critique. Wiley-Blackwell, Oxford.
- Chen, L., Mi, X., Comita, L.S., Zhang, L., Ren, H., Ma, K., 2010. Community-level consequences of density dependence and habitat association in a subtropical broadleaved forest. Ecol. Lett. 13, 695–704.
- Chen, L., Swenson, N.G., Ji, N., Mi, X., Ren, H., Guo, L., Ma, K., 2019. Differential soil fungus accumulation and density dependence of trees in a subtropical forest. Science 366, 124–128.
- Chesson, P., 2000. Mechanisms of maintenance of species diversity. Annu. Rev. Ecol. Syst. 31, 343–366.
- Comita, L.S., Hubbell, S.P., 2009. Local neighborhood and species' shade tolerance influence survival in a diverse seedling bank. Ecology 90, 328–334.
- Comita, L.S., Muller-Landau, H.C., Aguilar, S., Hubbell, S.P., 2010. Asymmetric density dependence shapes species abundances in a tropical tree community. Science 329, 330–332.
- Comita, L.S., Queenborough, S.A., Murphy, S.J., Eck, J.L., Xu, K., Krishnadas, M., Beckman, N., Zhu, Y., 2014. Testing predictions of the Janzen-Connell hypothesis: A meta-analysis of experimental evidence for distance- and density-dependent seed and seedling survival. J. Ecol. 102, 845–856.
- Connell, J.H., 1971. On the role of natural enemies in preventing competitive exclusion in some marine animals and in rain forest trees. Dynamics of populations 298, 312.
- Dai, N., Kenji, S., Akiko, S., 2002. Seedling establishment of deciduous trees in various topographic positions. J. Veg. Sci. 13, 35–44.
- Davis, M.A., Wrage, K.J., Reich, P.B., 1998. Competition between tree seedlings and herbaceous vegetation: Support for a theory of resource supply and demand. J. Ecol. 86, 652–661.
- Fang, X., Shen, G., Yang, Q., Liu, H., Ma, Z., Deane, D.C., Wang, X., 2017. Habitat heterogeneity explains mosaics of evergreen and deciduous trees at local-scales in a subtropical evergreen broad-leaved forest. J. Veg. Sci. 28, 379–388.
- Forcier, L.K., 1975. Reproductive strategies and the co-occurrence of climax tree species. Science 189, 808–810.
- Gilbert, G.S., Harms, K.E., Hamill, D.N., Hubbell, S.P., 2001. Effects of seedling size, El Niño drought, seedling density, and distance to nearest conspecific adult on 6-year survival of Ocotea whitei seedlings in Panamá. Oecologia 127, 509–516.
- Grubb, P.J., 1977. The maintenance of species-richness in plant communities: the importance of the regeneration niche. Biol. Rev. 52, 107–145.
- Gurevitch, J.S., Scheiner, M., Fox, G.A., 2006. The ecology of plants. Sinauer Associates, Sunderland, Massachusetts, USA.
- Harms, K.E., Powers, J.S., Montgomery, R.A., 2004. Variation in small sapling density, understory cover, and resource availability in four neotropical forests. Biotropica 36, 40–51.

#### H. Liu et al.

- Hogan, K.P., 1986. Plant architecture, photosynthetic pesponses, and population biology in the Palms Socratea Durissima and Scheelea Zonensis on Barro Colorado Island, Panama (Physiological, Tropical). University of Illinois Urbana-Champaign, Illinois US, Ph.D.
- Hubbell, S.P., 2001. The unified neutral theory of biodiversity and biogeography. Princeton University Press, Princeton&Oxford.
- Hyatt, L.A., Rosenberg, M.S., Howard, T.G., Bole, G., Fang, W., Anastasia, J., Brown, K., Grella, R., Hinman, K., Kurdziel, J.P., Gurevitch, J., 2003. The distance dependence prediction of the Janzen-Connell hypothesis: a meta-analysis. Oikos 103, 590–602.
- Inman-Narahari, F., Ostertag, R., Hubbell, S.P., Giardina, C.P., Cordell, S., Sack, L., 2016. Density-dependent seedling mortality varies with light availability and species abundance in wet and dry Hawaiian forests. J. Ecol. 104, 773–780.
- Janzen, D.H., 1970. Herbivores and the number of tree species in tropical forests. Am. Nat. 104, 501–528.
- Johnson, D.J., Beaulieu, W.T., Bever, J.D., Clay, K., 2012. Conspecific negative density dependence and forest diversity. Science 336, 904–907.
- Johnson, D.J., Bourg, N.A., Howe, R., McShea, W.J., Wolf, A., Clay, K., 2014. Conspecific negative density-dependent mortality and the structure of temperate forests. Ecology 95, 2493–2503.
- Johnson, D.J., Condit, R., Hubbell, S.P., Comita, L.S., 2017. Abiotic niche partitioning and negative density dependence drive tree seedling survival in a tropical forest. Proceedings. Biological sciences 284, 20172210.
- Johnson, P.C., 2014. Extension of Nakagawa & Schielzeth's R<sup>2</sup> GLMM to random slopes models. Methods in Ecoloy and Evolution 5, 944–946.
- Jurinitz, C.F., Oliveira, A.A., Bruna, E.M., 2013. Abiotic and biotic influences on earlystage survival in two shade-tolerant tree species in Brazil's Atlantic forest. Biotropica 45, 728–736.
- Kitajima, K. & Fenner, M. (2000) Ecology of seedling regeneration. Seeds: the ecology of regeneration in plant communities (ed M. Fenner), pp. 331-359. CABI, Oxon & New York.
- Kuang, X., Zhu, K., Yuan, Z., Lin, F., Ye, J., Wang, X., Wang, Y., Hao, Z., 2017. Conspecific density dependence and community structure: Insights from 11 years of monitoring in an old-growth temperate forest in Northeast China. Ecol. Evol. 7, 5191–5200.
- Lentz, K.A., 1999. Effects of intraspecific competition and nutrient supply on the endangered northeastern bulrush, scirpus ancistrochaetus schuyler (Cyperaceae). The American Midland Naturalist 142, 47–54.
- Lewis, S.L., Tanner, E.V., 2000. Effects of above-and belowground competition on growth and survival of rain forest tree seedlings. Ecology 81, 2525–2538.
- Lin, F., Comita, L.S., Wang, X., Bai, X., Yuan, Z., Xing, D., Hao, Z., 2014. The contribution of understory light availability and biotic neighborhood to seedling survival in secondary versus old-growth temperate forest. Plant Ecol. 215, 795–807.
- Lin, L., Comita, L.S., Zheng, Z., Cao, M., 2012. Seasonal differentiation in densitydependent seedling survival in a tropical rain forest. J. Ecol. 100, 905–914.
- Liu, H., Ma, Z., Yang, Q., Fang, X., Lin, Q., Zong, Y., Wang, X., 2017. Relationships between established seedling survival and growth in evergreen broad-leaved forest in Tiantong. Biodiversity Science 25, 11–22.
- Liu, H., Shen, G., Ma, Z., Yang, Q., Xia, J., Fang, X., Wang, X., 2016. Conspecific leaf litter-mediated effect of conspecific adult neighborhood on early-stage seedling survival in a subtropical forest. Sci. Rep. 6, 37830.
- Liu, X., Liang, M., Etienne, R.S., Wang, Y., Staehelin, C., Yu, S., 2012. Experimental evidence for a phylogenetic Janzen-Connell effect in a subtropical forest. Ecol. Lett. 15, 111–118.

- Liu, Y., Fang, S., Chesson, P., He, F., 2015. The effect of soil-borne pathogens depends on the abundance of host tree species. Nat. Commun. 6, 10017.
- Mangan, S.A., Schnitzer, S.A., Herre, E.A., Mack, K.M., Valencia, M.C., Sanchez, E.I., Bever, J.D., 2010. Negative plant-soil feedback predicts tree-species relative abundance in a tropical forest. Nature 466, 752–755.
- Murphy, S.J., Wiegand, T., Comita, L.S., 2017. Distance-dependent seedling mortality and long-term spacing dynamics in a neotropical forest community. Ecol. Lett. 20, 1469–1478.
- Nakagawa, S., Schielzeth, H., 2013. A general and simple method for obtaining R2 from generalized linear mixed-effects models. Methods Ecol. Evol. 4, 133–142.
- Norghauer, J.M., Malcolm, J.R., Zimmerman, B.L., 2006. Juvenile mortality and attacks by a specialist herbivore increase with conspecific adult basal area of Amazonian Swietenia macrophylla (Meliaceae). J. Trop. Ecol. 22, 451–460.
- Packer, A., Clay, K., 2000. Soil pathogens and spatial patterns of seedling mortality in a temperate tree. Nature 404, 278–281.
- Paine, C.E.T., Harms, K.E., Schnitzer, S.A., Carson, W.P., 2008. Weak competition among tropical tree seedlings: Implications for species coexistence. Biotropica 40, 432–440.
- Peay, K.G., 2018. Timing of mutualist arrival has a greater effect on *Pinus muricata* seedling growth than interspecific competition. J. Ecol. 106, 514–523.
- Peters, H.A., 2003. Neighbour-regulated mortality: the influence of positive and negative density dependence on tree populations in species-rich tropical forests. Ecol. Lett. 6, 757–765.
- Song, Y., Wang, X., 1995. Vegetation and flora of Tiantong national forest park. Zhejiang province Shanghai Scientific Documentary Press, Shanghai.
- Taylor, K.M., Aarssen, L.W., 1989. Neighbor effects in mast year seedlings of Acer saccharum. Am. J. Bot. 76, 546–554.
- Terborgh, J., 2012. Enemies maintain hyperdiverse tropical forests. Am. Nat. 179, 303–314.
- Tilman, D., 1982. Resource competition and community structure. Princeton University Press, Princeton, NJ.
- Tomé, M., Tomé, J.A., Araújo, M.C., Pereira, J.S., 1994. Intraspecific competition in irrigated and fertilized eucalypt plantations. For. Ecol. Manage. 69, 211–218.
- Wang, P., Shu, M., Mou, P., Weiner, J., 2018. Fine root responses to temporal nutrient heterogeneity and competition in seedlings of two tree species with different rooting strategies. Ecol. Evol. 8, 3367–3375.
- Wang, X., Kent, M., Fang, X., 2007. Evergreen broad-leaved forest in Eastern China: Its ecology and conservation and the importance of resprouting in forest restoration. For. Ecol. Manage. 245, 76–87.
- Webb, C.O., Peart, D.R., 1999. Seedling densitydependence promotes coexistence of Bornean rain forest trees. Ecology 80, 2006–2017.
- Weiner, J., Thomas, S.C., 1986. Size Variability and Competition in Plant Monocultures. Oikos 47, 211–222.
- Wills, C., 1996. Safety in diversity. New Sci. 149, 38-42.
- Willson, M.F., Traveset, A., 2000. The ecology of seed dispersal. Seeds: The ecology of regeneration in plant communities. CABI, Oxon &, New York, pp. 85–110.
- Wu, G., Jiang, S., Liu, H., Zhu, S., Zhou, D., Zhang, Y., Luo, Q., Li, J., 2018. Early direct competition does not determine the community structure in a desert riparian forest. Sci. Rep. 8, 4531.
- Xing, J., 2015. A precisely measuring method and the spatio-temporal heterogeneity of forest light regimes. East China Normal University, Shanghai, China, Master.
- Yang, Q., Ma, Z., Xie, Y., Zhang, Z., Wang, Z., Liu, H., Li, P., Zhang, N., Wang, D., Yang, H., Fang, X., Yan, E., Wang, X., 2011. Community structure and species composition of an evergreen broad-leaved forest in Tiantong's 20 ha dynamic plot Zhejiang Province, eastern China. Biodiversity Science 19, 215–223.